

Deforestation around the World and its impact on soil fertility and quality

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Introduction:

Deforestation means the conversion of forest to other land use or the long-term reduction of the tree canopy cover below the minimum 10 percent threshold. It includes areas of forest converted to agriculture, pasture, water reservoirs and urban areas. It also includes areas where, for example, the impact of disturbance, over-utilization or changing environmental conditions affects the forest to an extent that it cannot sustain a tree cover above the 10 percent threshold (**FAO, 2001**).

Deforestation defined broadly can include not only conversion to non-forest, but also degradation that reduces forest quality - the density and structure of the trees, the ecological services supplied, the biomass of plants and animals, the species diversity and the genetic diversity.

Forest is determined both by the presence of trees and the absence of other predominant land uses. The trees should be able to reach a minimum height of 5 m in situ. It includes areas with bamboo and palms provided that height and canopy cover criteria are met, forest roads, firebreaks and other small open areas; forest in national parks, nature reserves and other protected areas, such as those of specific scientific, historical, cultural or spiritual interest, windbreaks, shelterbelts and corridors of trees with an area of more than 0.5 ha and width of more than 20 m, plantations primarily used for forestry and protection purposes, such as rubberwood plantations and cork oak stands.

Forest Lost across the World:

Sl. No.	Country	Lost (mh) in 2017	% lost from 2016
1.	World	29.4	-
2.	Colombia	0.42	46
3.	Democratic Republic of Congo	1.5	6.2
4.	Indonesia	decrease	88

(Carrington *et al.*, 2018)

Marginal increase in India's forest:

Total forest cover in 2017	7,08,273 sq km
Total forest cover in 2015	7,01,495 sq km
Net increased	6778 sq km

Forest cover has decreased the most in the north-eastern states:

States	Forest lost
Mizoram	531 sq km
Nagaland	450 sq km
Arunachal Pradesh	190 sq km
Tripura	164 sq km
Meghalaya	116 sq km

(Indian state of forests report, 2017, Forest survey of India)

Important agents leading to deforestation:

The "agents" are those individuals, corporations, government agencies, or development projects that clear the forests as opposed to the forces that motivate them.

Table 1: Important agents leading to deforestation

Agents	Links to deforestation, degradation and fragmentation
commercial farmers	clear the forest to plant commercial cash crops, sometimes displace slash-and-burn farmers who then move to the forest
slash-and-burn farmers	clear forest to grow subsistence and cash crops
Livestock herders	intensification of herding activities can lead to deforestation
commercial tree planters	clear mostly forest fallow or previously logged forests to establish plantations to supply fibre to

	the pulp and paper industry
firewood collectors	intensification of firewood collection can lead to deforestation
mining and petroleum industrialists	roads and seismic lines provide access to other land users, localized deforestation related to their operations
land settlement planners	relocation of people into forested areas as well as settlement projects displacing local people who then move to the forest
infrastructure developers	new access for other land users from road and highway construction through forested areas, flooding by hydroelectric dams

Table 2: Important agent groups and their regions

Region	Main agents
Africa	1. slash-and-burn farmers 2. commercial farmers 3. loggers 4. livestock herders 5. refugee and civil disturbances
Asia-Oceania	1. commercial farmers 2. slash-and-burn farmers 3. loggers 4. commercial tree planters 5. infrastructure developers
Latin America and Caribbean	1. slash-and-burn farmers 2. cattle ranchers 3. commercial farmers 4. loggers 5. infrastructure developers

The causes of deforestation are many and varied.

Direct Causes:

Natural causes: Hurricanes, Natural fires, Pests, Floods

Human activity: Agricultural expansion, Cattle ranching, Logging, Mining and oil extraction, Construction of dams Roads

Underlying Causes:

Market failures

Unpriced forest goods and services, Monopolies and monopolistic forces

Mistaken policy interventions

Wrong incentives, Regulatory mechanisms, Government investment

Governance weakness

Concentration of land ownership, Weak or non-existent ownership, land tenure arrangements, Illegal activities and corruption.

Broader socioeconomic and political causes

Population growth and density, Economic growth, Distribution of economic and political power, Excessive consumption, Toxification, Global Warming, War.

Advantages of Deforestation:

It produces lumber and charcoal, offers job opportunities, removes diseased trees, allows for more agricultural land, provides residential land.

Disadvantages of Deforestation:

It harms wildlife and destroys their habitat, leaves a scar on the environment, affects the water cycle, contributes to the problem of climate change, causes soil erosion.

Impacts of deforestation on soil fertility and soil quality:

Soil fertility refers to the ability of a soil to sustain agricultural plant growth, i.e. the ability to supply essential plant nutrients and water in adequate amounts and proportions for plant growth and reproduction and the absence of toxic substances which may inhibit plant growth.

Soil quality means the ability of a soil to function within ecosystem boundaries to support healthy plants and animals, maintain or enhance air and water quality and support human health and habitation.

Soil fertility and quality consists of three components viz. physical, chemical and biological.

Table 3: Selected physical attributes for the cultivated soil and forest soils. Db = bulk density, n = porosity, AW = available water

Property	VerticHapludolls			FluventicCalciudolls		
	Cultivated Soil	Forest soil	% Change	Cultivated Soil	Forest soil	% Change
Sand (g kg ⁻¹)	123.3	193.3	36	186.6	243.3	23.3
Silt (g kg ⁻¹)	386.6	380.0	2.0	436.6	426.6	2.3
Clay (g kg ⁻¹)	490.	426.6	15	376.6	330.0	14.1
Db (g cm ⁻¹)	1.63	1.36	19.8	1.36	1.25	8.8
n (%)	37.0	47.0	21.3	52.5	58.3	9.9
AW(%)	10	18	44.4	15	17	11.8

(Rezapour&Alipour, 2017)

The soils under natural forest showed higher sand and lower clay and silt fractions than that of soils under cultivation. This may be due to the migration down the clay fraction into soil profiles (**Rezapour, 2014**). Such finding is evidenced by the higher clay content in the 50–110 cm soil layer in the forest land (depth distribution data not shown). Surely, the forest canopy shades the soil surface which can: (a) lead to enhanced infiltration of rainfall water into the soil and (b) serves as a vapour barrier and suppresses evaporation against moisture losses from the soil (**Khresatef**

al., 2008). Such conditions could result in more migration of clay fraction and subsequently its accumulation in the deeper horizons.

A pattern of increased bulk density (a rise of 14–20%) and decreased porosity values (a drop of 10– 22%) was observed for the samples of cultivated sites compared to those of the adjoining forest soils. Although, neither of these values is a limiting factor for plant rootgrowth, higher bulk density of the deforested site could result in a lower soil quality as compaction could occur and result in non-favourable aeration (**Khresat et al., 2008, Hajabbasi et al., 1997**). The loss of organic matter and the effects of compaction of machinery and human traffic through agricultural practices (such as ploughing and harvesting crops) can be the most important subject to increasing bulk density in the cultivated soil as compared to the forest land.

Table 4: Runoff content and Soil erosion of adjacent Natural forest, Garden and Cultivated land.

Variables	Forest	Garden	Cultivated land
Runoff (L/M ³)	16.39	12.17	12.95
Soil erosion (g/m ²)	11.47	15.5	18.21

(**Kavian et al., 2013**)

Results of rainfall simulation indicated that change in land use affected the amount of runoff content. Runoff content at 20 cm depth was lower in the garden area (12.17 ± 1.25 L/M²) and cultivated area (12.95 ± 0.58 L/M²) than in the natural forest (16.39 ± 1.23 L/M²) area. These results are discordant with those of **Martinez-Mena et al., (2008)**, **Girmay et al., (2009)** and **Fang et al., (2012)** who found that the change from forest to another land use led to a significant increase in total runoff. The highest antecedent soil moisture in samples from the natural forest compared with the other land uses is the main reason for the higher runoff in the natural forest. Higher initial soil water contents may cause reduction in soil pore space and consequently a reduction in permeability, and hence, the soil infiltrability is reduced under higher soil water contents.

Soil erosion was affected by land use (Table 4). As the results show, soil erosion of the garden and cultivated soil was 1.351 and 1.587 times higher than the amount of the natural forest soil, respectively. This result is consistent with the studies of **Martinez-Mena et al., (2008)**, **Girmay et al., (2009)** and **Fang et al., (2012)**.

Table 5: Selected Chemical attributes for the cultivated soil and forest soils

Property	VerticHapludolls	FluventicCalciudolls
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	Cultivated Soil	Forest soil	% Change	Cultivated Soil	Forest soil	% Change
pH	7.16	6.53	9.6	6.96	6.6	5.5
EC (dS m ⁻¹)	0.65	1.59	59.1	0.61	0.63	3.2
CEC (cmol(+) kg ⁻¹)	23.0	28.5	19.3	18.5	23.2	20.3
C:N	11.68	10.97	6.5	20.77	11.32	83.5

(Rezapour&Alipour, 2017)

An increasing pattern in pH values was observed in the soils under cultivation (a rise of 5–14%) than the forest soils. However, pH values were in an optimal range (6–7.5) for ‘general plant growth and microbial activity’ following deforestation and cultivation. The observed trend in soil pH under cultivation could be attributed to the mixing of surface soil with subsurface soil containing carbonates and basic compounds through tillage practices **(Rezapour and Samadi, 2012)**. Another possible explanation may be attributed to the high rate of organic matter in the forest land that produces organic acids and provides H⁺-ions in the soil solution thereby reducing soil pH **(Angassa et al., 2012)**.

Regarding the values of EC, both forest and cultivated soils were not saline and hence general plant growth and microbial activity did not seem to be limited due to the presence of salts. However, the soil EC value was significantly higher in the majority of the cultivated soils compared to soils under forest. The observed high level of salinity in the forest land may be explained by the recycling of basic compounds (e.g. Na and K) through tree-root uptake, translocation into plant shoots, and re-entry into the soil with litter fall and decomposition **(Rezapour, 2014)**.

The C:N ratio was low, 10–20, in both soils under forest and indicating that those soils are in a mineralized state with significant implication on organic matter cycling and nutrient release **(Angassa, 2012)**. Decomposition is slowed when the C:N ratio is high (>30:1) and more rapid when the C:N ratio is low (<20:1) **(Tisdale, 1993)**. The cultivated soil manifested a considerable increase in the values of C:N ratio, ranging from 3% to 84%, compared with the adjoining forest soils which is consistent with the results of Abbas et al., 2007 who observed high C:N ratios for cropland compared to soil under forest.

The values of CEC, as a favorable indicator of the fertility and productivity of soils, were medium to high categories, ranging from 18.5 to 23.0 cmolc kg⁻¹ for the cultivated soils and from 23.2 to 28.5 cmolc kg⁻¹ for the adjacent forest soils. The conversion of forest land to the cropland induced a drop of 19.3% (VH) to 20.3% (FC) in the amount of CEC. Typically, this pattern matched the distributive trend of organic carbon, suggesting the significant contribution of organic matter to CEC.

Table 6: Mass (kg ha⁻¹) of total nutrients in the forest floor

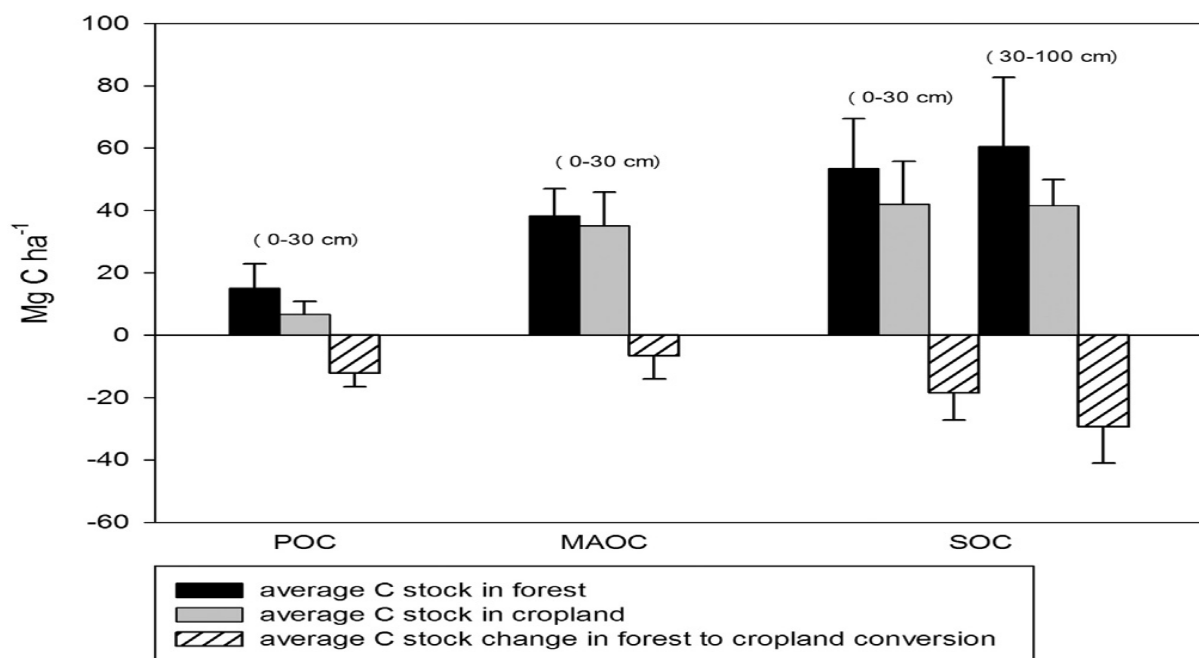
Study area	St-pere		
	Burned	Cut	Control
N	1060b	1870a	2130a
P	81a	120a	120a
K	105b	210a	182a
Ca	330b	670a	400b

Mg	44b	92a	57b
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Simardet *al.*, 2001

Nutrient element loss is a common effect of both clear-cutting and wildfire. Nitrogen is often the most limiting nutrient in boreal ecosystems and can be lost in large quantities during wildfire. This study showed a significant loss in the total mass of N and K in the forest floors of the St-Père burned stands. Nitrogen, which volatilizes at 200°C, is easily lost during wildfires, which can achieve temperatures of 300–700°C. Other less volatile nutrients, such as base cations and phosphorus are subject to loss through ash convection, wind and water erosion, especially more mobile ions, such as K⁺, which are lost through leaching. The youngest study area (St. Père) was the only area in which scarification and planting were included as part of the cut treatment. The literature suggests that site preparation commonly increases the loss of nutrients from disturbed ecosystems. However, mass and concentration of total nutrients in the forest floor of the cut stands did not drop significantly below levels in controls in any study area, suggesting that clearcutting and scarification did not cause significant nutrient loss.

Fig 1: Average stocks of total soil organic C (SOC), particulate organic C (POC) and mineral associated organic C (MAOC) in forest and in croplands, and average C stock changes in forest to cropland conversion (Villarino *et al.*, 2016)



Carbon losses account for nearly 19 Mg C ha⁻¹ of SOC, composed by 12 Mg C ha⁻¹ of POC and 7 Mg C ha⁻¹ of MAOC. SOC stock losses at 30–100 cm depth account for 30 Mg C ha⁻¹, and this value is around 11 Mg C ha⁻¹ higher than the SOC stock loss at 0–30 cm. In the forest, the observed SOC stock in the first meter of soil was 114 Mg C ha⁻¹, almost twice of the total C biomass. The average estimated SOC losses after 10 years of continuous cropping (48 Mg C ha⁻¹) was slightly lower than C stored in biomass (59 Mg C ha⁻¹) under forest, (Gasparriet *al.*,

2008). Therefore, deforestation led to a loss of approximately 104Mg C ha⁻¹ (56Mg ha⁻¹ in total biomass plus 48 Mg ha⁻¹ in SOC stock). The most widespread deforestation method in the study area consists in land clearing with heavy bulldozers, burning the remaining vegetation and then plowing down the residues (**Boletta et al., 2006**). This practice produces strong changes in soil, such as surface exposure to precipitation, wind and solar radiation, aeration and temperature increases, root removal, and aboveground biomass and charcoal incorporation. These changes severely affect SOC dynamics in different ways. Surface exposure to climatic conditions, aeration and temperature increases, and root removal could deplete SOC stocks, either by mineralization increases or by water and wind erosion. However, aboveground biomass and charcoal incorporation could have the opposite effect and this could explain the higher SOC stocks during the first 2–3 years of cropping. Soil organic C, POC, and MOAC (0–5 and 0–30 depths) losses under longer periods of cropping are probably due to decreases in the above ground net primary production (ANPP) in croplands (**Volante et al., 2012**), increases in the mineralization rate due to higher temperature and aeration, and wind erosion (**Rojas et al., 2013**). POC is considered a key soil quality indicator (**Haynes, 2005**), it is very likely that high losses under cropping are indicating strong soil degradation in the region.

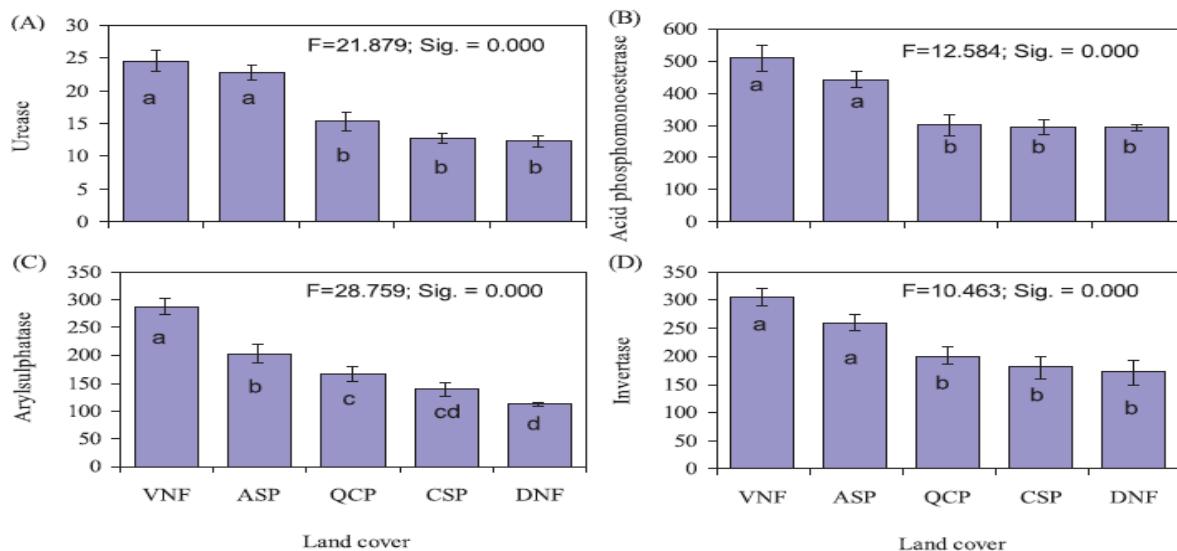
Table 7: Soil microbial and biochemical properties analyzed (Kooch et al., 2018)

Soil features	Virgin natural forest (VNF)	Degraded natural forest (DNF)
Microbial biomass C (mg kg ⁻¹)	685.50 a	453.62 c
Microbial biomass N (mg kg ⁻¹)	65.74 a	41.52 d
Microbial biomass P (mg kg ⁻¹)	81.31 a	37.93 e
Particulate organic C (mg kg ⁻¹)	3.15 b	1.21 c
Particulate organic N (mg kg ⁻¹)	0.50 a	0.34 b
Dissolved organic C (mg kg ⁻¹)	68.96 b	39.21 c
Dissolved organic N (mg kg ⁻¹)	39.21 a	26.60 b

Soil MBN and MBP were different among the land covers, and the natural forest (65.74 and 81.31 mg kg⁻¹) had the highest values but the least were found in the degraded forests (41.52 and 37.93 mg kg⁻¹), respectively. The enhanced levels of soil MBC, MBN and MBP under natural forest and also rehabilitated area could be explained by the higher nutrient concentrations and soil organic matter as compared to degraded sites (**Aponte et al., 2013**). Thus, the higher microbial biomasses (C, N

and P) under natural forest as compared to other sites are mainly attributable to the greater availability of organic matter in these forest sites, which is in agreement with the findings of **Wang and Wang(2007)** in San Menjiang Forest. DOC and DON release is often controlled by different factors. DOC is the primary form of C that is transported from forest floor to mineral soils. In general, the leaching of freshly fallen litter and the decomposition of forest floor organic matter are thought to be the major sources of DOC and DON in forest soils (**Kooch and Bayranvand, 2017**).

Fig 2: Enzyme activity(**Koochet al., 2018**)



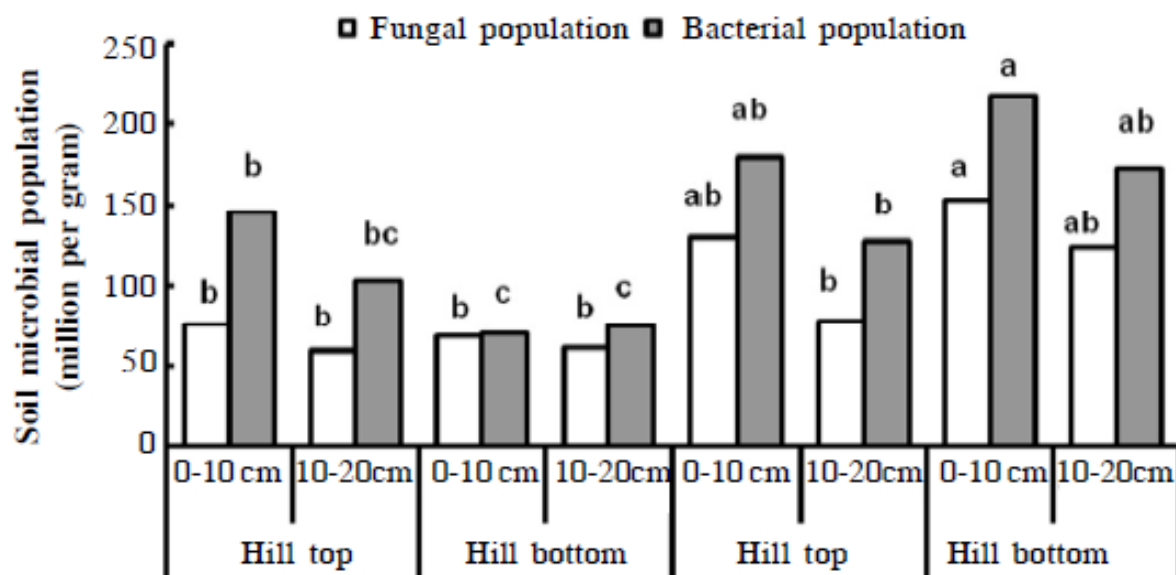
Soil activities were assigned to Urease ($\mu\text{g NH}_4^+-\text{N g}^{-1} \text{ h}^{-1}$); Acid phosphomonoesterase ($\mu\text{g PNP g}^{-1} \text{ h}^{-1}$); Arylsulfatase ($\mu\text{g PNP g}^{-1} \text{ h}^{-1}$); Invertase ($\mu\text{g Glucose g}^{-1} \text{ 3 h}^{-1}$). Virgin natural forest (VNF), Alnus subcordata C.A.M. plantation (ASP), Quercus castaneifolia C.A. Mey plantation (QCP), Cupressus sempervirens var. horizontalis plantation (CSP) and degraded natural forest (DNF).

Enzyme activities may be more quickly influenced by changes in forest management, which might be useful as an early indicator of biological changes (**Hu et al., 2006**). The activities of some of them like urease, invertase, phosphatases, and arylsulfatase, which are respectively involved in N, C, P, and S cycles, are more sensitive to land cover changes (**Raiesi and Beheshti, 2014**). The results obtained in this study show that the tree species in the ecosystem influenced the enzyme activities in the topsoil (Raiesi and Beheshti, 2014). The urease, acid phosphomonoesterase, arylsulfatase and invertase activity of virgin natural forest with native tree species (dominated by Carpinus betulus and Parrotia persica) was significantly higher than that of the Quercus castaneifolia and Cupressus sempervirens monoculture plantations with a negative effect on deforestation. This

suggested that a natural broad-leaved forest could have a better effect in promoting soil enzyme activity compared with pure plantations (**Pang et al., 2009**) of Quercus castaneifolia and Cupressus sempervirens as well as degraded sites. Parallel to virgin natural forest, after 30 years plantation of Alnus subcordata, the biomass on

the afforest soil surface increased and returned more litter to the soil, thus increasing urease, acid phosphomonoesterase and invertase enzyme activity. This emphasized the importance of long-term monitoring in order to adequately evaluate and understand the impacts of *Alnus subcordata* afforestation on degraded forestland soil properties.

Fig 3: Soil fungal and bacterial populations in deforested land and natural forest of Baroitoli Forest Beat, Chittagong South Forest Division. (Miah *et al.*, 2014)



Both fungal and bacterial populations were significantly ($p \leq 0.05$) lower in deforested land compared to adjacent natural forest in both the surfaces (0-10 cm and 10-20 cm depth) on top and bottom hill position in the study area (Fig. 3). In deforested land, fungal population ranged from 59 to 76 million g^{-1} dry soils while in natural forest it was 78 to 153 million g^{-1} dry soil. In deforested land, the bacterial population ranged from 70 to 146 million g^{-1} dry soil but in natural forest it was from 127 to 218 million g^{-1} dry soil. In land with the presence of forest, a definite ecological condition sets in with respect to light, moisture, drainage, aeration, organic food materials and both fine and coarse roots, in which a distinct type of fauna survives. Forest clearing and conversion to other land uses often result in a low level of microbial population (Jha *et al.*, 1992) and enzymatic activity (Salam *et al.*, 1998) due to changes in soil microclimate. In the present investigation, both fungi and bacterial populations were found higher in upper layer (0-10 cm) than in the lower layer (10-20 cm) at both the hill position of natural forest and deforested land. Cheng *et al.*, 1992 showed comparatively lower fungal and bacterial population with increased depth of soil profile from deforested hill slope of China, the similar tropical region of the present study.

Summary & conclusion:

The soils under natural forest showed higher sand and lower clay and silt fraction's than that of soils under cultivation. A pattern of increased bulk density and decreased porosity values was observed for the samples of cultivated sites compared to those of the adjoining forest soils. The rainfall runoff experiments indicate that runoff content of the natural forest soils was higher than the garden and cultivated land soils. Garden soil erosion and cultivated land was higher than the forest. pH values was high under cultivation than the forest soils. The cultivated soil manifested a considerable increase in the values of C:N ratio with the adjoining forest soils. The conversion of forest land to the cropland induced a drop in the amount of CEC. Soil organic C, POC, and MOAC losses under longer periods of cropping are probably due to decreases in the above ground net primary production (ANPP), increases in the mineralization rate due to higher temperature and aeration, and wind erosion. The microbial community *i.e.*, the fungal and bacterial population was also significantly lower in both surfaces (0-10 cm and 10-20 cm) of hill positions in the deforested land compared to natural forest. Microbial/biochemical indicators showed perceptible deterioration in the topsoil due to deforestation. In general, deforestation provides some advantages to human but it deteriorates soil fertility and quality significantly. So we need to stop deforestation and search for new way for our mankind.

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